



The Natural Background Activity Concentration of (^{226}Ra , ^{232}Th and ^{40}K) and the Annual Effective Dose from Different Water Sources Consumption in Phosphate Polluted Area

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THE present study is a part of series studies carried out on Nile valley phosphate mining areas, Egypt in order to gather information about radionuclides mobility and distribution and access the associated radiological health risk. This study was carried out on different water sources samples collected from Al-Mahamid, Upper Egypt where the study showed that the activity concentrations of ^{226}Ra and ^{232}Th are close to each other for all water sources samples and the activity concentration of ^{40}K is higher than that of ^{226}Ra and ^{232}Th for all water sources samples. These data are compared with those obtained for different countries. The estimated annual effective dose (D_{eff}) ranged from infants to adults were also examined. The total annual effective dose obtained for infant, children and adults are lower than that obtained by WHO and UNSCEAR. The health risks to members of the public due to exposure to low dose radiation from the various drinking water sources which is regarded as chronic risk of somatic or hereditary effects were estimated. The result shows that 56 out of 1,000,000 may suffer some form of cancer fatality and 20 out of 100,000 may suffer some hereditary effects.

Keywords: Phosphate / Radionuclides / Annual effective dose / WHO / UNSCEAR.

Introduction

Radionuclides are known to be associated with Phosphate rocks which contain relatively high concentrations of naturally occurring radioactive materials from the uranium and thorium decay series [1]. In April 2015, the International Atomic Energy Agency (IAEA) World Distribution of Uranium Deposits (UDEPO) estimated there were 13.8 million tons of uranium in phosphate rock deposits [2, 3 and 4]. There are no radiological controls on the operation of these industries or restrictions on how waste is discharged which relate to its radionuclide content [5]. When radioactivity released in the air, they can travel some distance depending upon many factors as wind speed and direction and altitude of the release. The products of airborne releases can be transported to humans by direct ingestion, inhalation or that deposited on the ground will find their way into plant and animal life and thereby into the food chain, or through deposition

of airborne into water which can reach humans either by direct ingestion or via the food chain. The radionuclide contributing significantly to the ingestion dose via consumption of water is radium [6]. Many salts of radium are soluble in water and therefore surface water may be enriched in radium and its progenies. ^{226}Ra is an earth alkaline element sharing the metabolic pathways of calcium in the human body [6 and 7]. Ingestion of natural radionuclides depends on the consumption rates of water and on the radionuclide concentrations. Ingested radionuclides are absorbed into the blood [8] and accumulates in specific tissues that they may damage. Of absorbed uranium, 66% is rapidly eliminated via urine while the rest is distributed and stored in the kidney (12-15%), bone (10-15%) and soft tissues. The rate of clearance of such radionuclide from the tissue or organ is dependent on the biological half-life. Higher concentration of radioactivity in environmental media can cause exposure risk to the general populace which may lead to radiation related sickness such as leukemia,

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Received : 1 / 5 / 2019 ; accepted : 16 / 6 / 2019

DOI : 10.21608/EJPHYSICS.2019.12416.1020

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cancer of bladder, kidney, testis and lungs [9, and 10]. World Health Organization (WHO) [11] recommended reference dose level of 0.1 mSv from one-year consumption of drinking water. This study was carried out on different water sources samples from phosphate polluted area in Al-Mahamid, Upper-Egypt in order to estimate the activity concentration of natural radionuclides ^{226}Ra , ^{232}Th and ^{40}K and assess the associated radiological health risk due to ingestion of such water.

Geographical aspects

Phosphate deposits in Egypt are part of the Middle East to the North African Phosphogenic Province of the Late Cretaceous-Palaeogene age. The occurrences are divided into three east-west trending facies belts, phosphorite of the northern facies belt spreading from Bahariya oasis to Sinai, phosphorite of the central facies belt which confined to three localities: (i) The Red Sea Coast from Safaga to the Quseir land-stretch. (ii) The Nile Valley between Idfu and Qena. (iii) The Western Desert on the Abu Tartur Plateau (New Valley area) and phosphorite of the southern facies belt [12]. The Nile Valley phosphate deposits extend between latitudes $25^{\circ} 30' - 26^{\circ} 30'$ and longitudes $32^{\circ} 30' - 33^{\circ} 30'$ on both sides of the Nile Valley [13]. Our study area is a part of the Nile Valley phosphate area in the area of Al-Mahamid, Upper Egypt ($25^{\circ} 6' 18.81''\text{N}$, $32^{\circ} 49' 15.53''\text{E}$ and $25^{\circ} 6' 4.11''\text{N}$, $32^{\circ} 49' 31.91''\text{E}$), where mining of phosphate ore is mostly by surface mining. The overburden is removed either by scraping or by drilling and blasting, depending on the nature of the rock. The phosphate bed is drilled, blasted and removed by trucks to the crushing plant, where it is crushed to less than 5 cm and screened, then attrition washed to remove the clayey fine fraction and the hard-siliceous coarse fraction the control of air pollution under these circumstances is very challenging [12 and 13]. During these operations, the dust spreads in varying degrees to cover neighboring areas [14].

Sampling and measurements

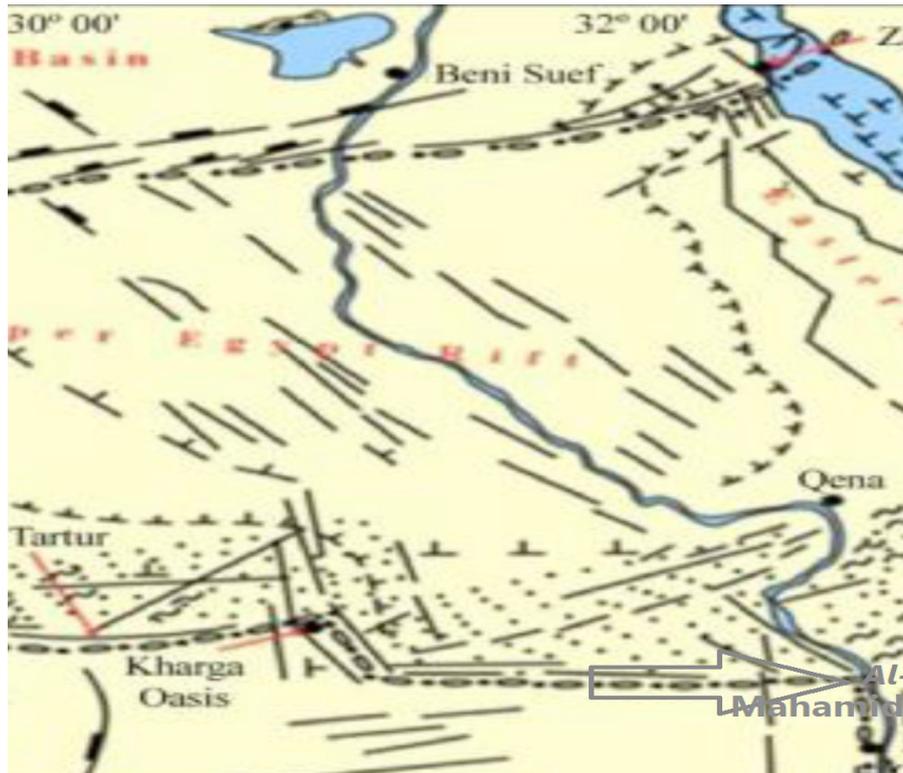
For the purposes of this study a number of different samples of water were collected from different sources and points to cover the total area under investigation (three water samples from three wells distributed randomly throughout the area, a canal water, Nile water as well as tap water from some houses around the study area). Samples were collected using polyethylene bottles 5-liters volume. Two samples were taken

from each position. The position was recorded using the global positioning system (GPS). The 5-liters volume of solution was concentrated to 1-liter volume [15]. Marinelli beakers (1-liters volume) were used as a measuring container. Each beaker was filled up to brim and a tight cap was pressed on so that the air was completely removed from it. The samples were stored for over 4 weeks to reach secular equilibrium before radiometric analysis. Low-level background gamma-ray spectrometer consists basically of 3×3 -inch NaI(Tl), S-1212-I model, with a 1024 microcomputer multichannel analyzer, 5510 ORTEC Norland was used to measure the specific activity of the samples. The applied detector has a peak gamma-ray efficiency of 1.2×10^{-5} at 1332 KeV, energy resolution of 7.5% at 662 KeV [16] and operation bias voltage 800-1000 V D.C. The detector was housed inside a massive lead shield to reduce the background radiation. The energy calibration was carried out by acquiring a spectrum from radioactive standards of known energies (^{137}Cs and ^{60}Co). For the efficiency calibration a multi-element standard of known activities was used. By assuming secular equilibrium in the uranium and thorium decay series, the ^{238}U and the ^{232}Th activities were determined indirectly via activities of their daughter. The nuclides chosen were ^{214}Bi (609.3, 1120.3 and 1764 KeV) and ^{214}Pb (351 KeV) for ^{238}U , ^{208}Tl (2614 KeV), ^{212}Pb (238 KeV) and ^{228}Ac (911 KeV) for ^{232}Th [17 and 18]. The specific activity of ^{40}K was determined directly by 1461 KeV photopeak. The background was measured frequently and subtracted from the net count for all measured samples. The activity concentrations of the natural radionuclides in the measured samples were computed using the following equation [19]:

$$A_s = \frac{N}{\epsilon P_r M} \quad (\text{e.M.Pr}) \quad (\text{Bq/Kg}) \quad (1)$$

Where N is the net counting rate of γ -ray (counts per second) corrected for background, ϵ the detector efficiency of the specific γ -ray, P_r the absolute transition probability of γ -decay and M the mass of the sample (Kg). The total annual effective dose due to intake of drinking water sources sampled for 3 different age groups (infants, children and adults) was estimated using the following formula [7, 10 and 20]: (2)

Where: D_{eff} is the annual effective dose equivalent from consumption of drinking water, A_r is the concentration of natural radionuclide in water (Bq/L), I_L is the consumption rate of water (L/y).



Map of the study area, Al-Mahamid, Upper Egypt.

TABLE 1. Dose conversion factors (mSv/Bq) for ingestion of radionuclides for members of the public [21].

| S/N | Radioisotopes | Infant < 1 year | Children 10 years | Adult > 17 years |
|-----|-------------------|-----------------|-------------------|------------------|
| 1 | ²²⁶ Ra | 5.7 E-06 | 8.0 E-07 | 2.8 E-07 |
| 2 | ²³² Th | 1.6 E-06 | 2.9 E-07 | 2.3 E-07 |
| 3 | ⁴⁰ K | 5.2 E-05 | 1.3 E-08 | 6.2 E-09 |

DC_R is the dose conversion factor (mSv/Bq) table (1). The water intake rates were 0.5 L/d and 1.0 L/d for infants (0-1 yrs) and children (10 yrs) respectively and 2 L/d for adults (≥ 17 yrs) [20]. The health risks to members of the public due to exposure to low dose radiation from the various drinking water sources which is regarded as chronic risk of somatic or hereditary effects were estimated [7].

$$\text{Fatality cancer risk} = D_{\text{eff}} \times C_{\text{RF}} \quad (3)$$

where C_{RF} is the cancer risk factor = 5.5×10^{-2}
 Lifetime fatality cancer risk to adult = fatality cancer risk $\times 70$ yrs (4)

$$\text{Severe hereditary effects} = D_{\text{eff}} \times H_{\text{EF}} \quad (5)$$

where H_{EF} is the hereditary effect factor = 0.2×10^{-2}

Lifetime hereditary effect in adult = Severe hereditary effects $\times 70$ yrs (6)

Results and Discussion

From table (2) and fig (1) the average activity concentrations of ²²⁶Ra, ²³²Th and ⁴⁰K were 1.35 ± 0.06 , 0.87 ± 0.04 , and 5.40 ± 0.20 Bq/L respectively for tap water (W.T) samples. The values of ²²⁶Ra ranged from 0.98 ± 0.05 to 3.04 ± 0.16 , ²³²Th ranged from 0.81 ± 0.04 to 1.84 ± 0.08 and ⁴⁰K ranged from 3.72 ± 0.13 to 10.47 ± 0.40 Bq/L for canal water (W.C) samples. For ground water (W.G) samples the values of ²²⁶Ra ranged from 1.99 ± 0.10 to 3.80 ± 0.18 , ²³²Th ranged from 1.30 ± 0.06 to 1.44 ± 0.07 and ⁴⁰K ranged from 5.07 ± 0.19 to 12.40 ± 0.45 Bq/L. The values of ²²⁶Ra ranged from 1.09 ± 0.05 to 2.56 ± 0.11 , ²³²Th 1.29 ± 0.06 to 1.46 ± 0.06 and ⁴⁰K from 8.01 ± 0.29 to 9.16 ± 0.33 Bq/L for Nile water (W.N) samples. The specific activity concentration of ⁴⁰K is greater than that of ²²⁶Ra and ²³²Th for all water

sources samples and the activity concentration of ^{226}Ra is more than that of ^{232}Th for almost water sources samples as ^{226}Ra is more soluble than ^{232}Th . Table 3 summarizes the concentration values of ^{226}Ra , ^{232}Th and ^{40}K which were obtained from the current study and other studies for other countries. As can be seen from the table 3 ^{226}Ra values for ground water of present study matches with the values reported by Isam, et al.[28]in Sweden and Godoy, et al.[31]in brazil and higher than the values reported by Ahmed [29] in Egypt (Qena) and (Safaga - Quseir) and lower than those reported by Salonen [24]in Finland. ^{232}Th values for ground, Nile, canal and tap water of present study close to each other and are matches with the values reported by Saqan, et al.[27] in Jordan. ^{40}K values of ground, Nile, canal and tap water for present study are lower than the values reported by Saqan, et al.[27]in Jordan. From table 4 and fig. 2 the annual effective dose for different age groups ranges from 0.037 to 0.103 $m\text{Sv/y}$ for infants, 0.0004 to 0.001 $m\text{Sv/y}$ for children and from 0.0004 to 0.0008 $m\text{Sv/y}$ for adult for canal water samples. In ground water, it ranges from 0.052 $m\text{Sv/y}$ to 0.120 $m\text{Sv/y}$ for infants,

0.0008 to 0.001 $m\text{Sv/y}$ for children and 0.0007 to 0.001 $m\text{Sv/y}$ in adult. For Nile water, the total effective dose ranges from 0.078 to 0.090 $m\text{Sv/y}$ for infant, 0.0005 to 0.0009 for children and 0.0005 to 0.0008 $m\text{Sv/y}$ for adult. It can be observed that the radiation dose received by infants is higher than that received by children and adults for all water sources samples. The total annual effective dose obtained for infant, children and adults are lower than that obtained by WHO [10] and UNSCEAR [20]. The result showed that cancer risk for adults varies from 20×10^{-6} to 56×10^{-6} for all water sources samples and Lifetime fatality cancer risk varies between 1×10^{-3} and 4×10^{-3} . Whereas the hereditary effect to adult per year varied from 0.728×10^{-6} to 2.036×10^{-6} and the lifetime hereditary effects in adult varies from 51×10^{-6} to 142×10^{-6} . the result shows that 56 out of 1,000,000 may suffer some form of cancer fatality and 20 out of 100,000 may suffer some hereditary effects. The United States Environmental protection Agency (USEPA) recommended acceptable cancer fatality risk limit of 1.0×10^{-6} to 1.0×10^{-4} (i.e. 1 person out of 1,000,000 to 10,000 persons suffering from some form of cancer fatality) [7, 10 and 32].

TABLE 2. The activity concentration of ^{226}Ra , ^{232}Th and ^{40}K (Bq/L) for different water sources (W.T= Tap water, W.C= Canal water, W.G= Ground water and W.N= Nile water)samples.

| Sample ID | Latitude | Longitude | $\pm A(^{226}\text{Ra})$ | $\pm A(^{232}\text{Th})$ | $\pm A(^{40}\text{K})$ |
|------------------|---------------|---------------|--------------------------|--------------------------|------------------------|
| W.T ₁ | 25° 7'56.72"N | 32°47'11.17"E | 1.30 ± 0.05 | 0.85 ± 0.03 | 5.61 ± 0.20 |
| W.T ₂ | 25° 7'56.99"N | 32°47'1.42"E | 1.40 ± 0.07 | 0.90 ± 0.04 | 5.20 ± 0.19 |
| W.C ₁ | 25° 7'58.67"N | 32°46'46.05"E | 2.77 ± 0.12 | 1.40 ± 0.06 | 6.06 ± 0.22 |
| W.C ₂ | 25° 7'56.09"N | 32°46'43.24"E | 2.11 ± 0.09 | 1.53 ± 0.07 | 4.92 ± 0.18 |
| W.C ₃ | 25° 7'52.99"N | 32°46'38.36"E | 1.46 ± 0.07 | 1.84 ± 0.08 | 10.28 ± 0.37 |
| W.C ₄ | 25° 7'41.22"N | 32°46'25.59"E | 0.98 ± 0.05 | 0.81 ± 0.04 | 6.01 ± 0.22 |
| W.C ₅ | 25° 7'47.56"N | 32°47'23.60"E | 3.04 ± 0.16 | 1.03 ± 0.06 | 10.47 ± 0.40 |
| W.C ₆ | 25° 7'24.45"N | 32°47'57.89"E | 2.54 ± 0.11 | 1.20 ± 0.05 | 8.84 ± 0.32 |
| W.C ₇ | 25° 6'22.25"N | 32°49'3.40"E | 1.06 ± 0.05 | 0.87 ± 0.04 | 3.72 ± 0.13 |
| W.G ₁ | 25° 7'48.67"N | 32°46'34.69"E | 3.80 ± 0.18 | 1.30 ± 0.07 | 5.07 ± 0.19 |
| W.G ₂ | 25° 7'44.12"N | 32°46'29.44"E | 1.99 ± 0.10 | 1.31 ± 0.06 | 12.40 ± 0.45 |
| W.G ₃ | 25° 7'33.74"N | 32°48'43.29"E | 2.59 ± 0.11 | 1.44 ± 0.06 | 8.49 ± 0.31 |
| W.N ₁ | 25° 7'31.13"N | 32°46'17.90"E | 2.56 ± 0.11 | 1.46 ± 0.06 | 9.16 ± 0.33 |
| W.N ₂ | 25° 6'58.46"N | 32°48'4.38"E | 1.63 ± 0.08 | 1.29 ± 0.06 | 8.01 ± 0.29 |
| W.N ₃ | 25° 6'20.28"N | 32°48'59.67"E | 1.09 ± 0.05 | 1.30 ± 0.06 | 8.21 ± 0.30 |

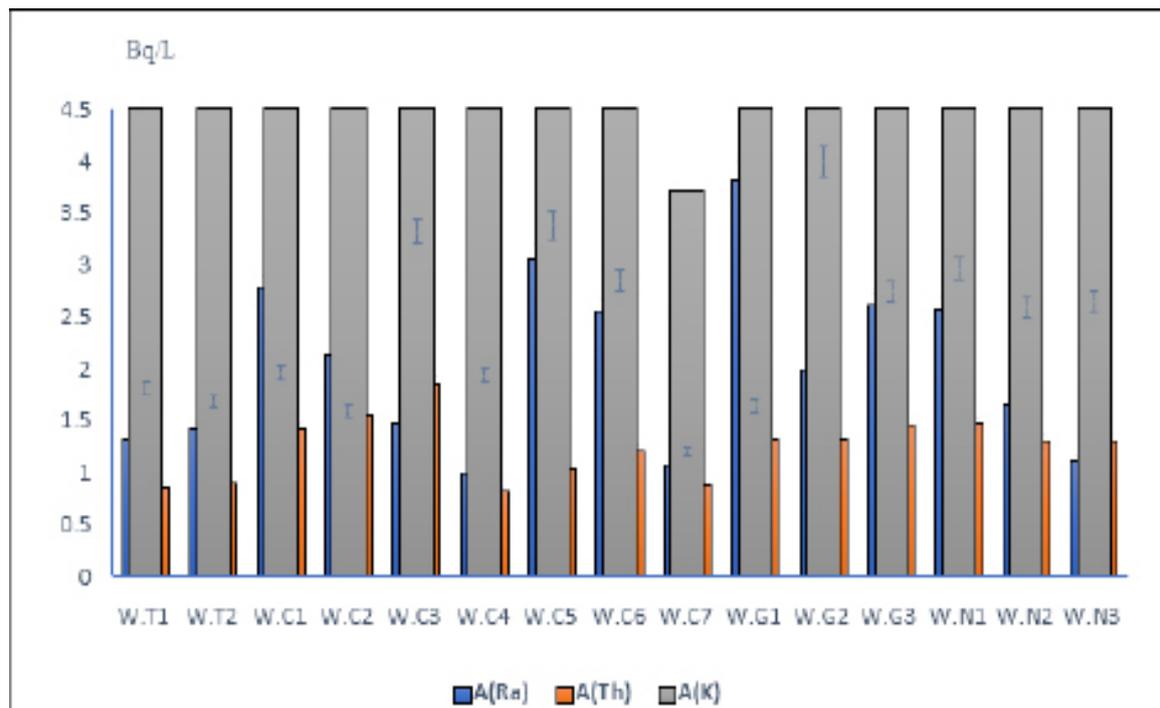


Fig. 1. The activity concentration of ²²⁶Ra, ²³²Th and ⁴⁰K (Bq/L) for different water sources (W.T= Tap water, W.C= Canal water, W.G= Ground water and W.N= Nile water)samples.

TABLE 3. The activity concentration of ²²⁶Ra, ²³²Th and ⁴⁰K (Bq/L) for water samples (Bq/L) in present study in comparison with other countries.

| Country | Water type | Activity concentration Bq/L | | | Ref. |
|-----------------------|-----------------------|-----------------------------|-----------|-----------|-------------------------------|
| | | Ra-226 | Th-232 | K-40 | |
| Current study | ground water | 1.99-3.8 | 1.3-1.44 | 5.07-12.4 | present study |
| | Nile water | 1.09-2.56 | 1.29-1.46 | 8.01-8.16 | present study |
| | canal water | 0.98-3.04 | 0.81-1.84 | 3.72-10.5 | present study |
| | tap water | 1.30-1.4 | 0.85-0.9 | 5.2-5.61 | present study |
| Egypt | natural water | 1.6-11.1 | 0.21-0.79 | 9.7-23 | El-Arabi, et al., 2006 [22] |
| Spain | natural water | 0.02-4 | | | Soto, et al., 1988 [23] |
| Finland | ground water | 0.01-49 | | | Salonen, 1994 [24] |
| Italy | d r i n k i n g water | 0.0002-1.2 | | | Sgorbati and Forte.1997 [25] |
| China | ground water | Max 0.93 | | | Weihai, Z., et al., 2001 [26] |
| Jordan | hot springs | 3.8-6.8 | 1.42-2.37 | 23.2-34.8 | Saqan, et al., 2001 [27] |
| Sweden | ground water | 0.016-4.9 | | | Isam., et al., 2002 [28] |
| Qena , Egypt | ground water | Mean 0.08 | Mean 0.04 | | Ahmed, 2004 [29] |
| Egypt (Safaga-Quseir) | ground water | Mean 0.1 | Mean 0.05 | | |
| U.S.A | mineral water | Max. 20 | | | Kitto, et al., 2005 [30] |
| brazil | ground water | 0.01 - 3.79 | | | Godoy, et al., 2006 [31] |

TABLE 4. The total annual effective dose, Cancerrisk to Adult, LifetimeFatality, Heredity effects to adult, hereditary effectsfor different water (W.T= Tap water, W.C= Canal water, W.G= Ground water and W.N= Nile water) samples.

| Sample ID | Total annual effective dose (mSv/y) | | | Fatality Cancerrisk to Adult per year $\times 10^{-6}$ | Lifetime Fatality cancerrisk | Severe Heredity effects to adult per year $\times 10^{-6}$ | Estimated lifetime hereditary effects |
|------------------|-------------------------------------|---------------|-----------------|--|------------------------------|--|---------------------------------------|
| | Infant ≤ 1 | Children = 10 | Adult ≥ 17 | | | | |
| W.T ₁ | 0.055 | 0.0005 | 0.0004 | 23.88 | 0.002 | 0.868 | 0.000061 |
| W.T ₂ | 0.051 | 0.0005 | 0.0005 | 25.36 | 0.002 | 0.922 | 0.000065 |
| W.C ₁ | 0.061 | 0.0010 | 0.0008 | 45.62 | 0.003 | 1.659 | 0.000116 |
| W.C ₂ | 0.049 | 0.0008 | 0.0007 | 39.13 | 0.003 | 1.423 | 0.000100 |
| W.C ₃ | 0.100 | 0.0007 | 0.0007 | 35.97 | 0.003 | 1.308 | 0.000092 |
| W.C ₄ | 0.058 | 0.0004 | 0.0004 | 20.01 | 0.001 | 0.728 | 0.000051 |
| W.C ₅ | 0.103 | 0.0010 | 0.0008 | 46.35 | 0.003 | 1.685 | 0.000118 |
| W.C ₆ | 0.087 | 0.0009 | 0.0008 | 41.84 | 0.003 | 1.522 | 0.000107 |
| W.C ₇ | 0.037 | 0.0004 | 0.0004 | 20.82 | 0.001 | 0.757 | 0.000053 |
| W.G ₁ | 0.052 | 0.0013 | 0.0010 | 56.00 | 0.004 | 2.036 | 0.000143 |
| W.G ₂ | 0.120 | 0.0008 | 0.0007 | 37.62 | 0.003 | 1.368 | 0.000096 |
| W.G ₃ | 0.084 | 0.0010 | 0.0008 | 44.56 | 0.003 | 1.620 | 0.000113 |
| W.N ₁ | 0.090 | 0.0009 | 0.0008 | 44.54 | 0.003 | 1.620 | 0.000113 |
| W.N ₂ | 0.078 | 0.0007 | 0.0006 | 32.22 | 0.002 | 1.172 | 0.000082 |
| W.N ₃ | 0.079 | 0.0005 | 0.0005 | 26.36 | 0.002 | 0.958 | 0.000067 |

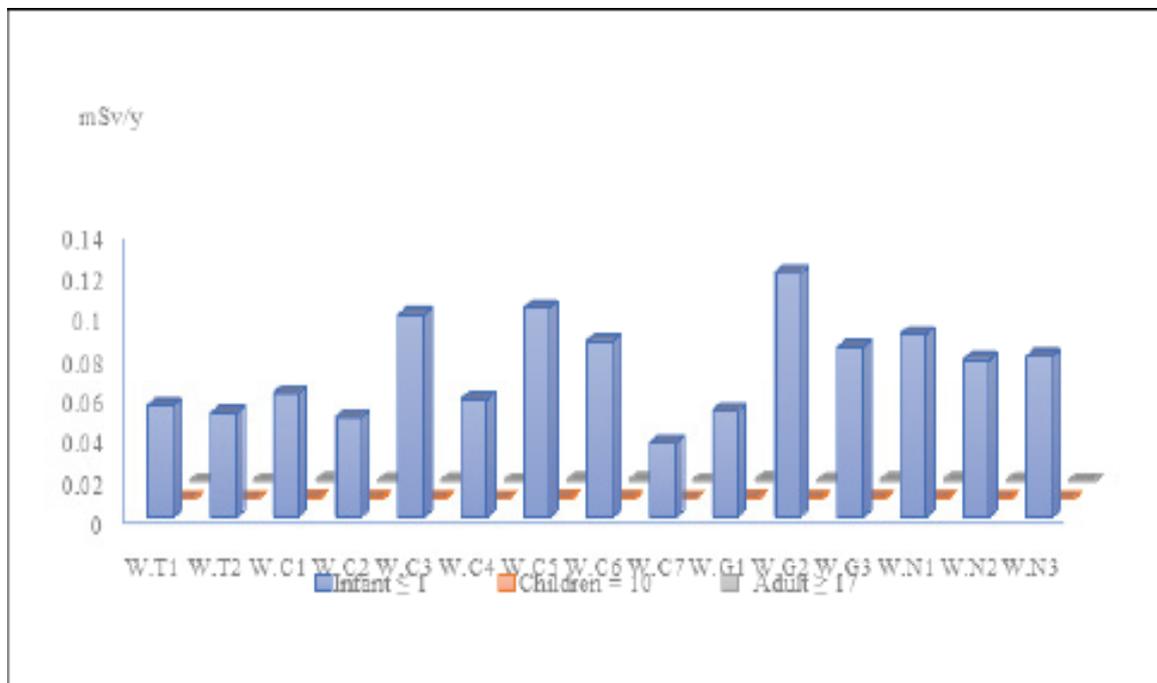


Fig. 2. The total annual effective dose (mSv/y) for different water sources (W.T= Tap water, W.C= Canal water, W.G= Ground water and W.N= Nile water) samples.

Conclusion

The distribution of natural radionuclides of ²²⁶Ra, ²³²Th, and ⁴⁰K were measured by gamma-ray spectrometry with NaI (TI) detectors for water sources samples in phosphate polluted area in Al-Mahamid, Upper Egypt. The results from this study indicate that the activity concentrations of ²²⁶Ra, ²³²Th, and ⁴⁰K in the studied water sources samples are found to be normal. The results of this study showed that the total annual effective dose obtained for infant, children and adults are below the limit set by WHO.

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تعتبر الدراسة الحالية واحدة من العديد من الدراسات التي تم إجرائها على مناطق تعدين خامات الفوسفات في مصر وغيرها من دول العلم وذلك لتقدير المخاطر الإشعاعية الناتجة من تعدين خامات الفوسفات نظرا لاحتوائها على تركيزات عالية من عنصر اليورانيوم المشع. حيث تم جمع عدد من عينات المياه من مصادر مختلفة (مياه جوفية، مياه النيل، مياه الترغ ومياه الحنفية) والتي يتم تناولها في البيئة المحيطة بمنطقة تعدين خام الفوسفات في منطقة المحاميد أدفو أسوان. حيث تم تقدير تركيز النويدات المشعة طبيعيا (الراديووم-22، الثوريوم-232 والبوتاسيوم-04) لهذه العينات وذلك باستخدام مطياف يوديد الصوديوم وتم حساب الجرعة السنوية الناتجة من تناول المياه من المصادر المختلفة والمخاطر الإشعاعية الناتجة لعدد من الفئات العمرية المختلفة (الرضع أقل من عام، الأطفال أقل من عشرة أعوام والشباب) وذلك طبقا للمعايير العيارية المحددة لمنظمة الصحة العالمية.